

Vegetation classification in south-western Australia's Mediterranean jarrah forest: new data, old units, and a conservation conundrum

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Abstract. Conservation reserve selection is guided by vegetation classification and mapping. New survey data and improvements in the availability of archived data through online data-sharing platforms enable updated classifications and the critique of existing conservation criteria. In the Northern Jarrah Forest Region of south-western Australia, percentage-based targets using 'forest ecosystem units' (15% of each unit) and the systematic conservation planning principles of 'comprehensiveness, adequacy and representativeness' underpin the State's reserve network. To assess the degree of community-level heterogeneity within the forest ecosystem units, new survey data for the forest (30 000 plots) were classified using a non-hierarchical clustering algorithm. Results were assigned to the National Vegetation Information System, and community groups defined at the Association level (Level V). Significant community level heterogeneity was found, including 15 communities in the dominant 'jarrah woodland' unit, and 13 in the 'shrub, herb and sedgelands' unit. Overall, this research highlights limitations in the current reserve system, including the influence of scale on percentage-based targets and 'representativeness'. A multi-scale approach to reserve selection, based on a quantitative, floristic, hierarchical classification system, would improve the level of scientific rigour underlying decision-making.

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Introduction

A primary goal of natural resource management is to balance development with conservation. In Australia and internationally, conservation reserve networks are key to achieving this (Cabeza and Moilanen 2001; Rondinini and Chiozza 2010; Kullberg and Moilanen 2014; Morelli *et al.* 2020; Stralberg *et al.* 2020). Vegetation classification and mapping underpin networks and provide a scientific basis for decision making. However, challenges remain. Baseline data are missing because of poor survey coverage, whereas incomplete and inconsistent classifications prevent assessment at suitable scales (McKenzie *et al.* 1996; Jennings *et al.* 2009; Franklin 2013; Peet and Roberts 2013). New data and the centralisation of datasets provide platforms for improvement and are occurring internationally (VegBank, BIEN, sPlot), nationally (TERN Aekos) and at State and Territory levels (e.g. BioNET (New South Wales), COREVeg (Queensland),

NatureMap (Western Australia), the Vegetation Site Database (Northern Territory); Benson 2008; Wisser and De Cáceres 2013; Chytrý *et al.* 2016; Faber-Langendoen *et al.* 2018; Gellie *et al.* 2018; Gibson 2018; Addicott and Laurance 2019; Bruelheide *et al.* 2019). Using these products to improve conservation assessment at multiple scales is key to improving decision-making.

Conservation decisions require targets (Bakker 2013). In Australia, this includes a continent-wide target of 17% of terrestrial ecosystems to be reserved, on the basis of the 2020 Convention on Biological Diversity's Aichi Target 11 (www.cbd.int/aichi-targets/target/11). More specifically, the National Reserve Strategy's (NRS) aim is a 'comprehensive, adequate and representative' (CAR) reserve system at the bioregional scale (JANIS 1997; Commonwealth of Australia 2010). The Interim Biogeographic Regionalisation for Australia (IBRA) and regional ecosystem units are the basis

for CAR targets (Commonwealth of Australia 2010). This includes the pre-1750 extent of vegetation and examples of at least 80% of the number of regional ecosystems in each IBRA region (comprehensiveness) or subregion (representativeness).

Although IBRA and regional ecosystem units provide a general framework for setting CAR targets, limitations are evident, particularly for defining and measuring representativeness. Representativeness is 'comprehensiveness considered at a finer scale' (Commonwealth of Australia 2010, p. 10); however, how IBRA subregions (rather than regions) are used to determine representativeness and detect finer-scale heterogeneity is not clear. The definition of regional ecosystems is also generic, with scale being a critical factor:

Ecosystems, or a unique unit such as a vegetation community mapped at appropriate scales, comprising a recognisable floristic composition in combination with a substrate (lithology, geology layers) and position in the landscape, and including their component biota [Commonwealth of Australia 2010, p. 65].

Work that tests how data collection and analysis methods affect the quantification of heterogeneity and representativeness will improve the scientific underpinning of the NRS.

State and Territory land and vegetation databases and map products provide the basis for this work. New South Wales BioNet collection includes over 800 maps of varying resolution, coverage and currency (Benson 2008). Queensland has 20 years of Regional Ecosystem Mapping (Neldner *et al.* 2019) and scale is recognised as important for detecting true floristic heterogeneity (Wardell-Johnson *et al.* 2007). Victoria's ecological vegetation classes (EVC) provide a range of units including mosaics, complexes, aggregates and generic wetland mapping units, which, in combination with bioregions, provide a bioregional conservation status (BRS) (databased in NatureKit 2.0). In Western Australia, state-wide mapping is at a 1 : 3 000 000 scale (Beard 1990; Shepherd *et al.* 2002; Beard *et al.* 2013). Broad floristic formations (forest ecosystem units) underpin the conservation reserve system, Regional Forest Agreement and Forest Management Plan (2014–2023; Department of Agriculture, Water and Environment 2019) (JANIS 1997; Commonwealth of Australia 1998; Government of Western Australia and the Commonwealth of Australia 1999; Conservation Commission of Western Australia 2013). Community-level mapping (1 : 5000–1 : 100 000) is available in some regions (Wardell-Johnson *et al.* 1989; Wardell-Johnson and Williams 1996; Mattiske and Havel 1997, 1998). This finer-scale mapping is used in development assessment, for determining offset targets, and informal reserve definition, but has not been fully incorporated into assessments of CAR for the conservation reserve system (JANIS 1997; Ecoscape 2002).

The Northern Jarrah Forest (NJF) region of south-western Australia provides a case study for investigating how community-scale data may influence CAR targets, including representativeness. A Mediterranean forest, the area has important commercial and conservation values, including as a major water catchment area for Perth (Petroni *et al.* 2010;

Hughes *et al.* 2012; Wardell-Johnson *et al.* 2015). Alcoa of Australia Ltd (Alcoa) have collected 30 000 plots of floristic data since the 1990s (see Trotter *et al.* 2018 for full details). The dataset covers a restricted area at high resolution, enabling community mapping and comparison to the broad floristic formations currently used to set CAR conservation targets in the forest.

Here, we use the dataset to assess how an updated classification informs mapping and conservation targets across scales in the NJF. Our aim is to describe and understand vegetation heterogeneity and its implications for decision-making. To do this, we perform the following:

- (1) Classify the data at three scales and use remotely sensed environmental attributes and modelled climate data to describe results. Landscape position (upland and plateau v. valley floors) is predicted to drive vegetation differences at broad scales, with finer patterns driven by local topography and soil.
- (2) Assign groups to the hierarchical National Vegetation Information System (NVIS).
- (3) Compare the groups derived from this analysis to the units currently used for mapping and decision-making, and assess how results inform the representativeness of the reserve system in the NJF.

Materials and methods

Study site

The NJF has a Mediterranean climate, with an average annual rainfall of 1300 mm in the west, dropping to 600 mm inland and to the east (Havel 1975b; Gentili 1989). A dry-sclerophyll system, it ranges from low woodlands in the north to open forests in the south (McKenzie 2008). The forest is significant globally, as Mediterranean ecosystems house a fifth of known vascular plant species in only 2% of the world's land area. They are also one of the biomes predicted to experience the greatest biodiversity change over the next 80 years, through both climate and human land-use pressures (Sala *et al.* 2000; Klausmeyer and Shaw 2009; Cox and Underwood 2011; Moreira *et al.* 2019). The NJF also occurs in the South-western Australia Global Biodiversity Hotspot (Myers *et al.* 2000; Hopper *et al.* 2016). Vegetation patterns are characterised by subdued topography, rainfall and regular fire incidence (Burrows *et al.* 2019). The forest also has a history of logging and mining (1834–current) that has considerably affected structure (Calver and Wardell-Johnston 2004; Wardell-Johnson *et al.* 2015). Floristically, the forest overstorey is largely dominated by two tree species, namely, *Eucalyptus marginata* (jarrah) and *Corymbia calophylla* (marri), but it has a complex and diverse understorey (3096 taxa; Dell *et al.* 1989; see <https://florabase.dpaw.wa.gov.au>). Understorey vegetation has been described as a multidimensional continuum with semi-discrete community types (Havel 1975a).

The climatic and hydrological regimes of the forest are shifting in response to global warming; rainfall has decreased by 16% and temperature increased 0.4°C since the mid-1970s (Bates *et al.* 2008; CSIRO, State of the Climate 2020, see csiro).

au/state-of-the-climate). This has led to decreased surface runoff (20–100%), falling water-tables and shifts from perennial to ephemeral streams (Petroni *et al.* 2010; Hughes *et al.* 2012; Zhang *et al.* 2016). High summer evapotranspiration offsets winter rainfall, which is predicted to further decrease under continuing climate change (CSIRO, State of the Climate 2020, see csiro.au/state-of-the-climate).

Floristic data

To meet statutory requirements, Alcoa has been collecting floristic data in the NJF for almost 30 years (1991–ongoing). Vegetation mapping now spans 432 km² with 31 000 plots on a 120-m² grid that follows ore exploration (Fig. 1). At each grid point, tree species were recorded in a 20-m radius and understory taxa in a 5-m radius. Cover-abundance (1–5) and stress (1–5) were scored (Mattiske 2012). Collected plant specimens were dried, fumigated and identified in accordance with the requirements of the Western Australian Herbarium (Mattiske 2012; see <https://florabase.dpaw.wa.gov.au>) and the conservation status of taxa were checked using current Government Gazette lists (see <https://florabase.dpaw.wa.gov.au>). Each plot in the dataset was also assigned a site

vegetation type, which is the community-scale unit used for mapping in the forest (Havel 1975a, 1975b). Eleven types were described across the study area.

Environmental data

Environmental variables were collected at each plot, and included soil type, topography, outcrop type and amount, dieback (*Phytophthora cinnamomi*) presence or absence and impact, years since fire, log debris and number of tree stumps. Additional environmental variables were generated in the System for Automated Geoscientific Analyses (SAGA, ver. 2.1.2; Conrad *et al.* 2015) using NASA's one arc-second (30 m) Shuttle Radar Topography Mission (SRTM) hydraulically enforced digital elevation model (DEM; Geoscience Australia 2011).

Topographic variables include distance to channel base network, which is calculated from an inverted DEM and negatively correlated with elevation (high valley-depth values indicate a larger distance from ridge lines, or lower valley position). Distance to valley depth is an inverted ridge-top distance. Curvature and convexity both measure the roundness or hollowness of a surface, with convexity

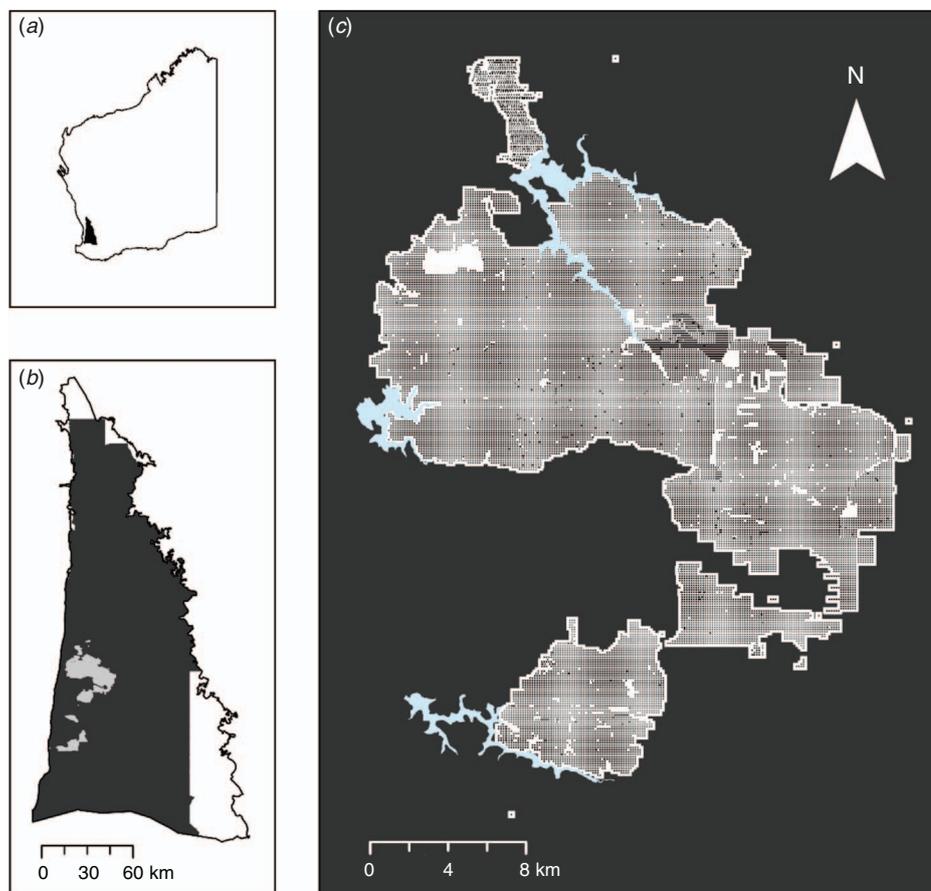


Fig. 1. Map of study area: (a) Western Australia (outline) and the Northern Jarrah Forest (black); (b) the study area (grey) within the mapped region (black) of the Northern Jarrah Forest (outline); (c) site data collection followed the 120-m drilling grid used by Alcoa for ore exploration (points). Dams and weirs (light blue).

indicating the number of convex shapes within a set radius, whereas curvature measures the convex or concavity of a slope. The SAGA wetness index (SWI) was used to represent flow across the study area. SWI is similar to the topographic wetness index (TWI; Beven and Kirkby 1979), but is based on a modified catchment-area calculation, which does not conceptualise flow as a thin film (Böhner *et al.* 2002).

Aspect in degrees was transformed to radians for further analysis (Roberts 1986). Northness (derived from $\cos(\text{aspect})$) has values close to 1 if the aspect is northward (or -1 if southward). Alternatively, eastness is a function of $\sin(\text{aspect})$, with values close to 1 representing east-facing slopes. Environmental data were tested for collinearity using the package *corrplot* (Wei *et al.* 2017). Where correlation was more than 0.5 (positive or negative), one variable from the pair was removed (Fig. S1 of the Supplementary material). Removal was based on previous research and expert opinion regarding which environmental driver would be the most valuable when interpreting vegetation patterns (e.g. Havel 2000).

Data cleaning

A floristic information management system was developed to prevent data corruption and data-entry error, and to ensure that future data additions fit system requirements (Trotter *et al.* 2018). This included automating name updates through Max (the Western Australian Herbarium platform for sharing plant name updates, see <https://www.dpaw.wa.gov.au/about-us/science-and-research/publications-resources/113-max-a-species-database-helper-for-windows>). In addition to the technical data-cleaning process, 337 species were excluded from the original Alcoa species list because of data-entry errors, subspecies updates and incorrect species identification. Singletons (a single species record) and doubletons (only two species records) were also removed from the dataset for this analysis (186 species). Annuals, biennials, deciduous taxa such as most geophytes, and alien taxa were removed because surveys were not conducted consistently in the flowering season. In total, 435 species remained in the modified dataset for this analysis. Floristic data were summarised in R (ver. 3.4.2, see <https://cran.r-project.org/bin/windows/base/old/3.4.2/>) and species richness and Shannon diversity were calculated.

Algorithm choice

Cluster analysis was performed using the non-hierarchical algorithm ALOC in PATN (ver. 3.12, see <https://patn.org/>; Belbin 1984, 1987; Belbin and McDonald 1993). Non-hierarchical algorithms have benefits over hierarchical ones, because they 'produce a single partition that optimises within-group homogeneity, instead of a hierarchical series of partitions optimizing the hierarchical attribution of objects to clusters' (Legendre and Legendre 2012, p. 349). The goal is to obtain a direct representation of the relationships among objects, rather than a summary of the hierarchical relationships that optimise fusions (Belbin 1987; Legendre and Legendre 2012). The potential to re-allocate objects to any group in the partition at each iteration is another advantage of non-hierarchical models, and computing memory requirements can be lower,

as an object-group, rather than object-object resemblance matrix is calculated (Legendre and Legendre 2012). Negatives include that users must pre-define the number of groups to be clustered and that the relationship between objects (usually shown by a dendrogram) can be difficult to visualise.

By calculating object-group, rather than object-object associations, the non-hierarchical algorithm ALOC is effective in clustering large datasets. The number of groups is pre-defined by the user and the number of iterations $n-1$ (n = sample size). The algorithm begins by allocating objects to a (suboptimal) set of seeds, starting with the first object in the dataset (although prior groups can also be used). It proceeds by taking each object, calculating the closest group and reallocating it to a closer group if more a homogenous option is found, until an overall optimal solution is reached (Belbin 1987). Overall, the algorithm is reasonably insensitive to seeds, and has been tested against hierarchical methods and shown to have a good recovery of clusters (Belbin and McDonald 1993).

Alternative non-hierarchical clustering algorithms that were investigated, but not selected for analysis, include k-means, partitioning around medoids (PAM), clustering large applications (CLARA), OPTPART and OPTSIL (MacQueen 1967; Kaufman and Rousseeuw 2009; Roberts 2015; Kassambara 2017). K-means is a commonly used unsupervised machine learning algorithm (Kassambara 2017). It classifies objects so that within cluster, similarity is maximised, and between-clusters it is minimised. There are several k-means algorithms; however, the standard one (Wong and Hartigan 1979) uses Euclidean distance, which is sensitive to outliers and not appropriate in data with many zeros, as is the case for many ecological datasets.

PAM (or k-medoids clustering) is less sensitive than is k-means to outliers and more likely to converge at a similar solution from different starting medoids (Kaufman and Rousseeuw 2009). However, PAM is considered suitable for at most 2000–3000 plots (R help file). CLARA is the recommended non-hierarchical alternative in R for larger datasets, but also uses Euclidean (or Manhattan) distance. Roberts (2015) stated an updated case for the use of non-hierarchical clustering with his algorithms OPTPART and OPTSIL. However, OPTSIL took 50 h of CPU time for 209 species and 424 sites (Intel I7 processor). Because this dataset has over 400 species and 30 000 sites, the computing time would have been large. Although high-performance computing provides solutions to this, fast desktop alternatives such as PATN (one simulation took an average of 20 min to complete) are a practical and effective alternative.

Cluster analysis and evaluation

For this study, a site by species matrix (31 087 sites; 435 species) was imported into PATN and the Data > Analysis option was selected. The pre-set menu of PATN splits analysis options by columns and rows. For rows, no association measure was selected, with non-hierarchical clustering (constant 2) and 99 final groups. Ninety-nine groups were selected to ensure that adequate community variation was captured, relative to the number of communities previously

described for the study area (11). For columns, a two-step association measure was selected, with 21 groups, to display site by species relationships (constant 2). All evaluations were selected (details below). The analysis was run, and results were exported to Excel for further processing.

PATN has a range of summary statistics that can be exported from the cluster analysis and used to visualise the results. This includes row (sites) and column (species) group statistics. Column summary data show the mean value of a species association to a group centroid, with the most frequent species having the highest values, whereas row data capture the group and distance to centroid that a site belongs to. Row statistics were used to produce an object–group dendrogram, to show how groups are related to one another, whereas species scores (column statistics) were used in non-metric multidimensional scaling (nMDS) to display species–group relationships. Visual assessment of the dendrogram was used to determine the number of clusters (groups) for further interpretation and evaluation.

Cluster evaluation was performed using the inbuilt evaluation tools of PATN (Belbin 1993; L. Belbin, unpubl. data). These include the Kruskal–Wallis statistic, principal-component correlation (PCC) and Monte-Carlo attributes in an ordination (MCAO). Kruskal–Wallis is a non-parametric (rank-based) test. In PATN, environmental variables were used to determine a significant difference among groups, with higher values being the most discriminating (Kruskal and Wallis 1952). PCC fits environmental variables to the cluster solution in ordination space, using multiple linear regressions (each variable is fitted independently) and provides an R^2 value for interpretation (Belbin 1993). MCAO tests the robustness of the PCC results. Using MCAO, environmental variables are re-allocated among objects (sites), the PCC linear regressions re-run 100 times and results compared with the true R^2 value. Results show the number of R^2 values that exceeded the true R^2 value (percentage permuted $R^2 > \text{actual } R^2$), with values higher than 5 of 100 indicating that a significant correlation is unlikely (PATN help file). Because of the limitations in PATN for the visualisation of non-hierarchical results in ordination space, the results of the clustering (species group statistics) were exported and displayed using nMDS (with the Bray–Curtis index) in R using the package *vegan* (ver. 2.3.1, J. Oksanen, F. G. Blanchet, M. Friendly, R. Kindt, P. Legendre, D. McGlenn, P. R. Minchin, R. B. O’Hara, G. L. Simpson, P. Solymos, M. H. H. Stevens, E. Szoecs, and H. Wagner, see <https://cran.r-project.org/web/packages/vegan/index.html>).

Typology and comparison to previous classifications

Following visual interpretation and cluster evaluation, 2, 11, 30 and 77 groups were selected for further analysis. The 30-group partition was chosen to represent community-level patterns, based on expert interpretation. Because no synoptic table or fidelity measure options are available in PATN, species frequency per group was calculated to show the dominant (indicator) species for each group (Table S1). A NVIS-style description was developed for each community group (McKenzie 2008; NVIS Technical Working Group

2017). Several attributes required for a NVIS description were inferred, because data were not collected in the field (stratum cover, dominant growth form and average height). Cover was based on species frequency per group, whereas the dominant growth form was assigned using FloraBase (see <https://florabase.dpaw.wa.gov.au/>) species descriptions. Strata were assigned on the basis of growth form, including upper layer (U): trees; mid-layer (M): shrubs, small trees and monocots; and ground layer (G): small shrubs, grasses, sedges, ferns, cycads and herbs. Boxplots and nMDS with the Bray–Curtis index were used to visualise species–group and environmental associations for the 2- and 30-group partitions in R. Summary statistics from the cluster analysis for the 30-group partition (mean distance from group centroids for each species) were used as input for the nMDS ordination.

An analysis of the difference between previously mapped units and the current results was performed using a confusion matrix of the frequency of site occurrence between the two typologies (see the R script file in the Supplementary material). Because 11 community types had been previously described across the study area, an 11-group partition of the classification was used for the comparison. Finally, confusion statistics (commission, omission and overall accuracy) were calculated for the updated typology.

Results

Classification

Overall, the classification identified two supergroups, representing the upland and riparian flora of the NJF (A and B, Fig. 2). These supergroups are heavily skewed; Supergroup B contains the majority of sites (95%) and species, with the largest subgroup (26) containing 404 species and 24 253 sites (Fig. 2, Table S1). Upland species dominate this supergroup (Table 1), reflecting their high abundance in the dataset (jarrah, marri and *Macrozamia riedleii*, interspersed with *Xanthorrhoea preissii*, *Bossiaea aquifolium*, *Allocasuarina fraseriana*, *Persoonia longifolia* and *Pteridium esculentum*). Species unique to the riparian supergroup (A, 1421 sites, Fig. 2) include *Eucalyptus patens*, *Taxandria linearifolia*, *Banksia littoralis* and *Hypocalymma angustifolium*, which are associated with valleys, stream banks and seasonally wet areas (see <https://florabase.dpaw.wa.gov.au/>). Key dominant species occur across both supergroups, notably jarrah, marri, *X. preissii*, *M. riedleii* and *P. esculentum* (Table 1, Table S3). Aspect, valley depth and convexity varied between the two supergroups, with riparian sites (A) being found lower in the landscape on less convex, north-east-facing slopes (Fig. 3). Species richness and the number of stumps (a measure of historic logging intensity) was also lower (Fig. 3).

At the community level (30 groups) *Eucalyptus patens*, *E. rudis* and *E. megacarpa* co-occur with the forest-wide dominants jarrah and marri in the upper stratum of riparian communities (Fig. S2, Table S3 of the Supplementary material). Mid-stratum species *Trymalium ledifolium* and *X. preissii* co-occur widely with both upland and riparian taxa. The major communities (26–30) were not well separated in ordination space, but 16 small groups

Table 1. NVIS hierarchical description for Supergroups A and B, from Class to Subassociation

Supergroup A includes the dominant riparian species in the forest (1421 sites), and Supergroup B contains the majority of species and sites (27 000 sites), including the common upland taxa. U, tree layer; M, shrub layer; G, ground species. Symbology and naming conventions follow the Australian Vegetation Attribute Manual (ver. 7.0, NVIS Technical Working Group 2017). \pm , the species may or may not occur in that vegetation type; \wedge , a second genus is co-dominant within a stratum (NVIS Technical Working Group 2017, p. 42)

Level	Description	Supergroup A	Supergroup B
I	CLASS	Tree	Tree
II	STRUCTURAL FORMATION	Woodland	Woodland
III	BROAD FLORISTIC FORMATION	<i>Eucalyptus</i> and <i>Corymbia</i> woodland	<i>Eucalyptus</i> and <i>Corymbia</i> woodland
IV	SUB-FORMATION	<i>Eucalyptus</i> and <i>Banksia</i> \wedge woodland; <i>Taxandria</i> and <i>Hypocalymma</i> \wedge ; <i>Pteridium</i>	<i>Eucalyptus</i> and <i>Corymbia</i> \wedge woodland; <i>Macrozamia</i> and <i>Banksia</i> \wedge ; <i>Hibbertia</i> and <i>Lomandra</i> \wedge
V	ASSOCIATION	U <i>Eucalyptus patens</i> , <i>Eucalyptus marginata</i> , <i>Corymbia calophylla</i> (tree); M <i>Taxandria linearifolia</i> , <i>Xanthorrhoea preissii</i> , <i>Banksia littoralis</i> (shrub, tree, monocot); G <i>Hypocalymma angustifolium</i> , <i>Macrozamia riedlei</i> , <i>Pteridium esculentum</i> (shrub, cycad, fern)	U <i>Eucalyptus marginata</i> , <i>Corymbia calophylla</i> (tree); M <i>Banksia grandis</i> , <i>Persoonia longifolia</i> , <i>Trymalium ledifolium</i> (shrub); G <i>Macrozamia riedlei</i> , <i>Lasiopetalum floribundum</i> , <i>Hibbertia commutata</i> (shrub, cycad)
VI	SUB-ASSOCIATION	U <i>Eucalyptus patens</i> , <i>Eucalyptus marginata</i> , <i>Corymbia calophylla</i> , \pm <i>Eucalyptus megacarpa</i> , \pm <i>Eucalyptus rudis</i> (tree); M <i>Taxandria linearifolia</i> , <i>Xanthorrhoea preissii</i> , <i>Banksia littoralis</i> , <i>Mirbelia dilatata</i> , <i>Melaleuca preissiana</i> (shrub, tree, monocot); G <i>Hypocalymma angustifolium</i> , <i>Macrozamia riedlei</i> , <i>Pteridium esculentum</i> , <i>Tetraria capillaris</i> , <i>Thomasia paniculata</i> (shrub, cycad, fern, sedge)	U <i>Eucalyptus marginata</i> , <i>Corymbia calophylla</i> , \pm <i>Allocasuarina fraseriana</i> (tree); M <i>Banksia grandis</i> , <i>Persoonia longifolia</i> , <i>Trymalium ledifolium</i> , <i>Xanthorrhoea preissii</i> (shrub, monocot); G <i>Macrozamia riedlei</i> , <i>Lasiopetalum floribundum</i> , <i>Hibbertia commutata</i> , <i>Boronia fastigiata</i> , <i>Lomandra sonderi</i> (shrub, cycad, herb)

(<20 sites) and uncommon taxa occurred as outliers, associated with restricted or distinct topographic features (Fig. 4). *Kunzea micrantha* (KUNMIC) and *Aotus cordifolia* (AOTCOR) are shrubs found in peaty soil, marshes and swamps, and *K. micrantha* forms a community with other riparian taxa including *Melaleuca preissiana*, *E. megacarpa* and *Lepidosperma squamatum*. The outlier *Philydrella drummondii* (PHDRU) is a perennial herb of freshwater swamps. Community 23 contains species associated with granite outcrops including the fern *Cheilanthes austrotenuifolia* (CHEAUS), *Neurachne alopecuroidea* (NEUALO) and *Hakea petiolaris* (HAKPET; FloraBase, see <https://florabase.dpaw.wa.gov.au/>). Several other outlier species are likely because of low records numbers or data-cleaning errors (e.g. *E. microcorys* and *Styloidium calcaratum*, an alien and annual species respectively; Fig. 4).

Overall, there were significant differences (MCAO < 5%) among the community groups in species richness, the number of stumps (a measure of logging disturbance), mean annual rainfall and aspect (northness; Table 2). Mean annual rainfall explained the most variation among groups ($R^2 = 0.32$, MCAO = 0%), whereas northness ($R^2 = 0.3$, MCAO = 0%) and years since fire ($R^2 = 0.28$, MCAO = 0%) were also significant (Table 2). At the supergroup level, Supergroup B was associated with higher wetness and valley depth values (a negative corollary of elevation) on reduced slopes (Fig. 3, Table S2). Both eastness and northness were positive for Supergroup A (0.32, 0.35; indicating north-east-facing slopes) and were slightly negative in Supergroup B

(-0.04, -0.03; indicating a very slight south-west-facing tendency; Fig. 3, Table S2).

Comparison to current mapping in the forest

Nine forest ecosystem units have been mapped for the study area (Fig. 5c). Broadly, they are comparable to the supergroups, there being geographic overlap between Supergroup A and the non-dominant forest ecosystem units (e.g. shrub, herb and sedgeland), whereas Supergroup B corresponds to the jarrah-north-west unit (Fig. 5c). The community (30 group) and 77 type classifications showed additional heterogeneity not captured by the forest ecosystem units (Table 3). For example, jarrah woodlands contains 15 communities and 50 types, whereas the shrub, herb and sedgeland unit incorporates 13 communities and 45 types (Table 3). Alternatively, the western wandoo woodland unit corresponds with a single community and only two swamp communities were described for the larger swamp unit.

At the community scale, previous mapping for the forest has been conducted using site vegetation types (SVT), with 11 being described in the study area (Table 4, Fig. 5a). There is some overlap between previous mapping and an 11-group cut of the current classification, with the majority of historic types (SVT) clustering into one major and two minor groups (overall accuracy = 45%, Table 4). Over half of the SVTs are found in the S and T groups (64% of plots); however, the current 11-group classification is even more unbalanced, with 95% of plots falling into Group 11 and seven groups

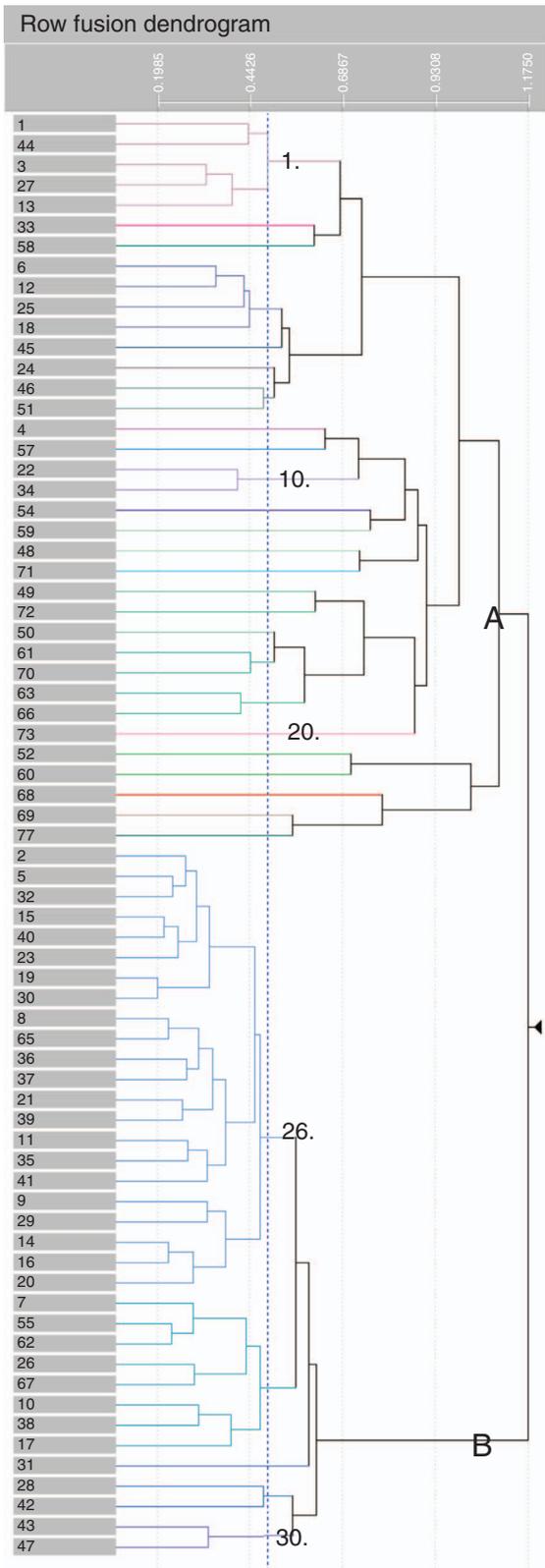


Fig. 2. UPGMA dendrogram of the classification showing two supergroups, 30 community groups and 77 types. There is a large degree of separation at the two-supergroup scale and more subtle differences among lower-level groups. Branch labels: supergroups (A and B) and communities (1–30), types (1–77).

having <10 quadrats each (Table 4). Group 1 contains the largest proportion of historic types associated with wet and riparian conditions (A, C, W; Table 4).

Discussion

Reserve systems are a key societal tool for balancing development with conservation (Pressey *et al.* 2007; Volenec and Dobson 2020). Reserve selection in Australia is based on systematic conservation planning (with the goal of being CAR), and a percentage area, regional approach is used to define comprehensiveness and representativeness (Commonwealth of Australia 2010; Kukkala and Moilanen 2013). This study used the classification of a large floristic database (~30 000 sites) to define vegetation types at several scales and test the representativeness of current units. Overall, two large supergroups (upland v. riparian) and 30 communities were described. The community classification includes rarer types that occur in topographically distinct (e.g. granite outcrops) or riparian areas, within a matrix of commonly occurring species across the uplands and plateaus. Tests of representativeness found that up to 15 communities occur within the regional-level units currently used for decision-making. Results demonstrated that community-level heterogeneity is not adequately captured by the conservation reserve system. Although it currently meets national CAR targets, additions are required to meet the criteria of representative at the community scale.

Explaining heterogeneity

The drivers of broad floristic patterns in the NJF are topography and climate, and the strong differences found between the riparian and upland supergroups in this study were to be expected on the basis of previous work (Havel 1975a, 1975b, 2000). Whereas key dominants occur across the full gradients of the forest, a distinct suit of species has been described for the valleys and creeks. Fire is also important, but its influence on community composition is less understood (Havel 2000; Burrows and Wardell-Johnson 2003; Williams *et al.* 2009; Enright *et al.* 2011, 2014; Wardell-Johnson *et al.* 2017; Burrows *et al.* 2019).

Additional factors that may influence patterns, but were not well investigated in this study, include soil pH, stochastic or random effects and biotic interactions. Laliberté *et al.* (2014) found that variation in plant species richness was almost entirely explained by soil pH, acting as an environmental filter between regional and local species pools (although this is not likely to be a driver in the NJF; Havel 1975a). In the kwongan heathlands of south-western Australia, multiple environmental factors did not explain differences among restored sites, with unmeasured stochastic or random effects being important (Riviera 2019). Finally, biotic interactions (e.g. below-ground facilitative relationships) may be important and are being incorporated into models where traditional variables cannot fully explain observations (Araújo and Luoto 2007; Wisz *et al.* 2013; le Roux *et al.* 2014; Pausas and Bond 2019).

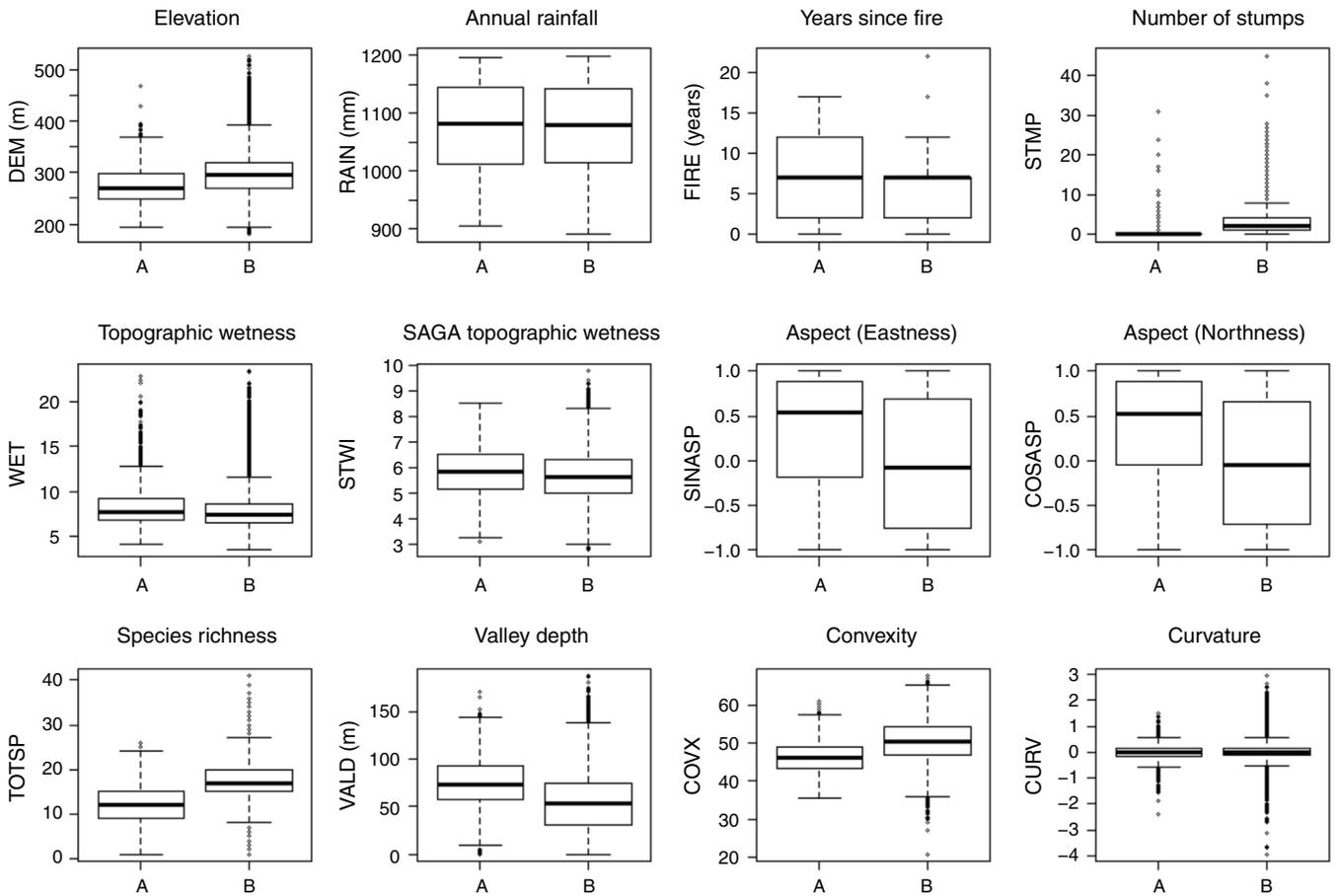


Fig. 3. Boxplots of environmental variables for the supergroup classification. Supergroup A (1421 sites) includes predominantly riparian species and is found lower in the landscape on less convex, north-east-facing slopes than is Supergroup B (27 000 sites), which is found on mid-slope to uplands and plateaux. DEM, digital elevation model; RAIN, average annual rainfall, 1960–2010; FIRE, years since fire; STMP, number of tree stumps in a 20-m radius; WET, topographic wetness index; STWI, SAGA wetness index; SINASP, eastness; COSASP, northness; TOTSP, species richness (number of species per plot); VALD, valley depth; CONVEX, convexity; CURV, curvature.

Methodological challenges

Defining clear groups at the community scale was difficult (Fig. 4). This was likely to be due to the nature of vegetation patterns and a lack of environmental gradients, exacerbated by the survey method (rapid sampling focused on upland areas of interest to mining). Previous work in the forest by J. Havel described the vegetation as a multidimensional continuum with semi-continuous patterns (Havel 1975a, 1975b). Havel also did not find clear community groups, despite surveying across larger environmental gradients than did this study and collecting site-based attributes (soil pH, N, K) for semi-supervised classification. Because patterns are not discrete, regression and other forms of gradient analysis may be more useful than the community concept for understanding vegetation in this system (Whittaker 1967; Ferrier *et al.* 2007; Austin 2013).

One type of regression that is well tested in ecological data is generalised dissimilarity modelling (GDM). GDM is a multivariate regression method that has been applied widely in conservation planning and biodiversity assessment (Mokany and Ferrier 2011; House *et al.* 2012; Molloy *et al.* 2016; Ferrier

et al. 2020; Mokany *et al.* 2020). Abiotic variables are used to model patterns in species turn-over at the community level and the following three methods are possible: predict then assemble, assemble then predict, or predict and assemble simultaneously (Ferrier 2002; Ferrier and Guisan 2006). Combining remotely sensed and modelled data (e.g. Sentinel-2, Landsat, the soil and landscape grid of Australia) with floristic information in GDM may enable community groups to be better separated across the upland and plateau areas of the NJF. Conversely, given the historic difficulties in defining vegetation communities in this system and findings from work in other Australian systems with gradual environmental gradients, supervised classification may not improve cluster separation (Havel 1975a, 1975b; Addicott and Laurance 2019).

Finally, decisions made during the definition of vegetation communities, including survey design and the type of field data collected (i.e. intuitive mapping *v.* full floristic plots), with the data analysis approach used, influence how community patterns are delineated (McKenzie 2008; Kent 2011). In systems such as the NJF, with many widespread, co-occurring species, subtle changes in less common species

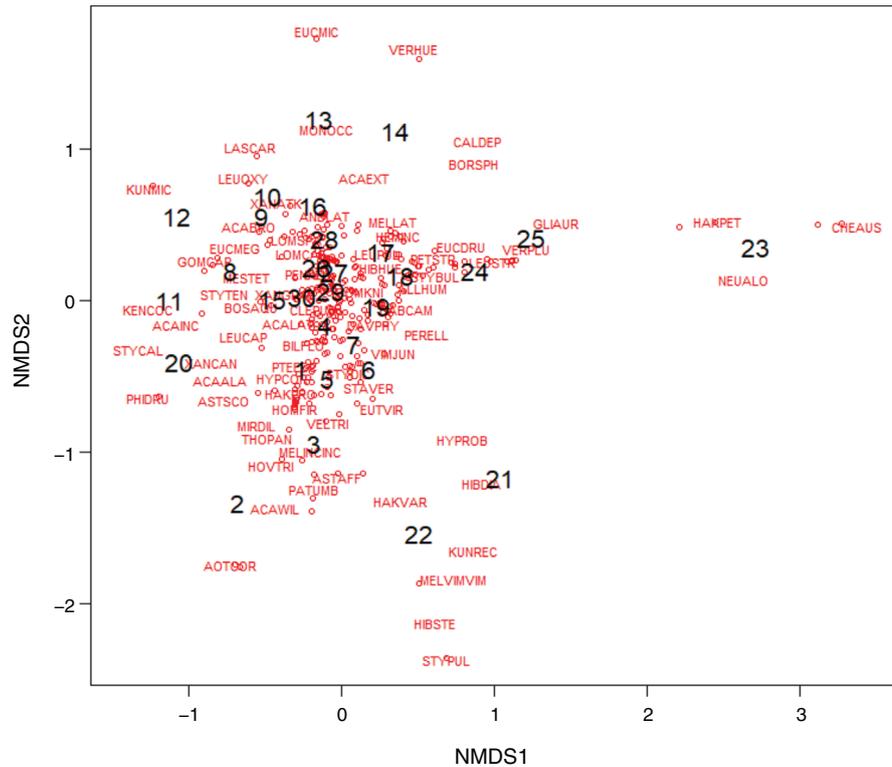


Fig. 4. nMDS with Bray–Curtis index for the community (30-group) classification, with selected species (red) and groups (black). Stress = 0.172. Red dots are species scores, with labels removed to aid overall readability.

Table 2. Evaluation statistics for the community (30-group) partition

Mean annual rainfall, northness and years since fire are significantly different among groups (MCAO < 5%), as is species richness and diversity. Principal-component correlation (PCC), Kruskal–Wallis (KW), Monte-Carlo attributes in an ordination (MCAO)

Environmental variable	PCC (R^2)	MCAO (%)	KW
Species richness	0.53	0	50.5
Number of stumps	0.44	0	60.4
Mean annual rainfall	0.32	0	38.4
Northness	0.30	0	59.0
Years since fire	0.28	0	48.2
Shannon diversity index	0.25	0	60.6
Distance to valley depth	0.21	0	49.2
Curvature	0.08	11	46.1
Convexity	0.05	22	56.1
Eastness	0.04	35	58.1
SAGA wetness index	0.04	45	51.7
Topographic wetness index	0.02	68	55.6
Longitudinal curvature	0.02	82	45.8

will be hard to detect using rapid-survey methods. Because these species are necessary to detect site differences (indicator species), unique community types will be difficult to describe (De Cáceres and Legendre 2009). These factors are likely to have had an impact on the results of this study and precipitate the creation of large units that do not accurately represent the complexity of a system.

A conservation conundrum

This study has highlighted the following two issues with the conservation reserve system in the NJF: (1) that the forest ecosystem units are not sufficient for the conservation of community-level heterogeneity; and (2) the limitations of a percentage-area approach.

Forest ecosystem units are the basis of a CAR reserve system in the NJF, despite being based largely on structural attributes and overlooking significant floristic heterogeneity (Table 3, Fig. 5c; JANIS 1997; Ecoscape 2002). Although work has been undertaken in the forest to account for community-scale patterns in decision-making (vegetation complexes, Fig. 5b), it is not legislated (Mattiske and Havel 1997, 1998). This study has provided additional evidence that vegetation heterogeneity is not sufficiently captured by the forest ecosystem units (Table 3, Fig. 5c). More broadly, heterogeneity is acknowledged as not being dealt with well in the conservation planning literature (Possingham et al. 2005; Kukkala and Moilanen 2013). Adaptive conservation management and policy is needed, with the flexibility to incorporate scientific developments (Reside et al. 2018). To ensure our reserve systems are CAR, data products and assessments at the community scale are required.

This study has also highlighted the limitations of using percentage-area targets for reserve system design. In addition to not sufficiently capturing fine-scale ecological heterogeneity, percentage targets can overlook ecological

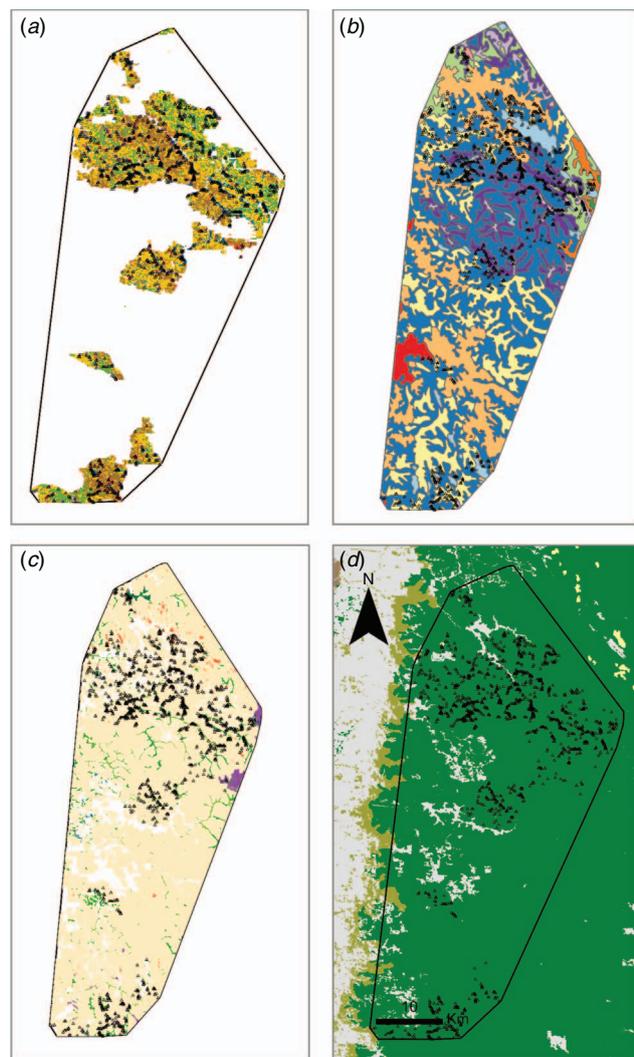


Fig. 5. Historic vegetation units available in the study area: (a) site vegetation types, (b) vegetation complexes, (c) forest ecosystem units; and (d) the National Vegetation Information System (NVIS), Level II – structural formation. Results from the supergroup classification are overlain – black triangles: riparian vegetation (Supergroup A, 1421 sites); and clear: upland and plateau Supergroup B (27 000 sites). Data for the forest ecosystem units shared with permission from the Western Australian Department of Biodiversity, Conservation and Attractions.

processes, climate change and other threats (Woinarski *et al.* 2007; Carwardine *et al.* 2009; Klein *et al.* 2009). These limits are being recognised and efforts made to improve the realism of conservation prioritisation models (Carwardine *et al.* 2009; Runge *et al.* 2016; Woodley *et al.* 2019). In the NJF, minimum reserve requirements have been met (15% of pre-1975 extent is reserved), but the impacts of climate change are measurable (16% decrease in average rainfall since 1970). Modelling the impacts of drying and changing fire regimes to identify refugia and incorporating these areas into the reserve network would provide a bridge from percentage-area targets to an adaptive

Table 3. The number of communities (30 groups) and types (77) in the dataset, for each forest ecosystem unit

Finer-scale mapping shows heterogeneity within units, indicating that representativeness is not captured at the regional scale

Forest Ecosystem Unit	30 Groups	77 Types
Bullich and Yate	3	11
Jarrah–North East	4	16
Jarrah–North West	7	>28
Jarrah–Sandy Basins	2	3
Jarrah woodland	15	50
Rocky outcrops	6	28
Shrub, herb, and sedgelands	13	45
Swamps	2	7
Western Wandoo woodland	1	3

management approach (Klein *et al.* 2009; Keppel *et al.* 2012, 2015; Kukkala and Moilanen 2013; McLaughlin *et al.* 2017; Morelli *et al.* 2020; Stralberg *et al.* 2020).

Conclusions

This study has highlighted that representativeness (a measure of heterogeneity at fine scales) is not well accounted for by the forest ecosystem units currently used for reserve selection in the NJF. On average, six community groups occur within each unit, with 13 being described in the shrub, herb and sedgelands unit and 15 in the jarrah woodland. The current reserve system of the NJF, although meeting its target of 15%, is not likely to capture an equal portion of each community group within these areas.

There are several means to improve measures of representativeness in both the NJF and national reserve systems. One is the incorporation of high-quality survey data collected at the optimal time of year (full floristics) to capture non-dominant species, which may be important indicators of unique community types. The second is the re-analysis of existing datasets to develop and test measures of representativeness. Critiquing current systems and defining limitations will provide a basis for adaptive management and prioritising improvements in reserve selection. The third is improvements in the pipeline from data collection and analysis to policymaking. Decisions made during the definition of vegetation communities, including survey design and the type of field data collected (i.e. intuitive mapping v. full floristic plots), with the data analysis approach used, influence how community patterns are delineated. Finally, a multi-scale approach to reserve selection, based on a quantitative, floristic, hierarchical classification system will improve the scientific rigour underlying decision-making. Although defined targets are practical for administration purposes, better incorporation of complexity will bring us closer to truly comprehensive, adequate and representative reserve systems.

Conflicts of interest

Sarah Luxton and Grant Wardell-Johnson are Associate Editors for *Australian Journal of Botany*. Despite this relationship, they did not at any stage have editor-level

Table 4. Confusion matrix for community-scale historic mapping (site vegetation types or SVT, A–W) v. an 11-group classification of the current data (Groups 1–11, ordered by relative frequency to the site vegetation types)

The updated typology does not separate the variation described by site vegetation types well, with high commission and omission errors, an overall accuracy of 45% and only three core groups delineated (1, 6 and 11). Above and below the diagonal shows when results from the two typologies are the same (i.e. how sites are distributed across the 'SVT' typology for each of the 11 'groups' and vice versa)

Groups	Site vegetation types											Total	Commission error (%)
	A	C	D	E	G	P	Q	R	S	T	W		
8	6	0	2	0	0	0	0	0	1	0	0	9	33
1	471	487	86	6	1	19	1	4	26	9	98	1208	60
6	4	1	76	2	52	12	0	6	6	1	2	162	53
3	0	0	0	0	0	0	0	0	0	0	1	1	100
10	0	0	0	0	26	0	0	0	1	0	0	27	3.7
4	0	1	0	0	0	0	0	0	0	0	0	1	100
5	2	0	0	0	1	0	0	0	2	1	0	6	100
7	0	0	1	0	0	0	0	0	0	0	0	1	100
11	104	251	897	61	94	5960	80	187	12 215	5985	1199	27 033	55
2	0	0	1	0	0	0	0	0	4	0	0	5	100
9	0	0	0	0	1	0	0	0	0	0	0	1	100
Total	587	740	1063	69	175	5991	81	197	12 255	5996	1300	28 454	
Omission error (%)	99	34	93	100	85	100	100	100	0.3	100	100		

access to this manuscript while in peer review, as is the standard practice when handling manuscripts submitted by an editor to this journal. *Australian Journal of Botany* encourages its editors to publish in the journal and they are kept totally separate from the decision-making process for their manuscripts. The authors have no further conflicts of interest to declare.

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